Assessing the Impacts of Environmental and Ecological Variables on the Performance of Fraser Sockeye Salmon Forecast

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**Abstract**

Fraser River sockeye salmon is one of the largest stock complexes in North America, supporting major commercial, recreational, and First Nations fisheries. Sockeye fisheries management relies on an annual pre-season forecast of adult returns. In this study, we developed a framework using Taylor diagram to facilitate evaluating pre-season forecast models annually and identify important external drivers for forecasting sockeye returns. The Taylor diagram provides a concise statistical summary and graphical framework that allows forecasts produced from a range of models to be compared to observations. Specifically, we incorporated five new covariates including sea surface temperature in the Gulf of Alaska and other salmon species abundance into the existing forecast models. Results revealed that new covariates improved the performance of all models. In addition, models selected a decade ago underperformed compared to recent retrospective analysis from 2009-2020. We recommend that forecast models be continuously evaluated in the face of increasing environmental change, and more environmental and ecological factors be explored and incorporated into forecast models for assessing for their impacts on the dynamics of Fraser sockeye.

**Key words**: Fraser sockeye, environmental covariates, *Oncorhynchus nerka*, pre-season forecast, salmon abundance, Taylor diagram

1. **Introduction**

Sockeye salmon (*Oncorhynchus nerka*) make significant economic, ecological, and social contributions throughout the Northeast Pacific basin. The sockeye salmon complex in the Fraser River, British Columbia, Canada, is one of the largest stock complexes in North America, which spawn in hundreds of distinct locations throughout the 220 000 km2 drainage basin (Figure 1; Burgner 1991; DFO 2018). This stock complex has supported major commercial, recreational, and First Nations marine and terminal fisheries (Ruggerone and Connors 2015), making the sustainable management of these sockeye salmon stocks imperative. The management system of these sockeye salmon consists of pre-season forecast of sockeye returns (a combination of sockeye to be caught by fisheries and those escaping to spawning ground), in-season management and post-season review. Pre-season forecasts are essential and have become mandatory for Fisheries and Oceans Canada (DFO) under a Pacific Salmon Treaty that initiated in 1985 (PSC 2022). Pre-season forecasts of Fraser sockeye returns are used to plan fishery opportunities by fisheries managers and the fishing industry, and feed to in-season assessment modeling (Haeseker et al. 2008; Michielsens and Cave 2019; DFO 2021). Additionally, pre-season forecast assists decision-making processes including international total catch allowance negotiation, allocation of stock assessment resources, and hatchery enhancement planning (Hawkshaw et al. 2020a, b).

Fraser sockeye salmon returns are highly variable not only due to the cyclic nature for some large stocks (with a dominant year preceded and followed by three years of low returns), but also unpredictable productivity at different life stages in both freshwater and marine environments (Ricker 1997; Akenhead et al. 2016a; Huang et al. 2021), which has made pre-season forecasts of sockeye returns challenging.

Despite the tremendous importance and challenges of pre-season Fraser sockeye forecasts, a rigorous evaluation of forecast performance in terms of both accuracy and uncertainty has not been completed. In particular, confidence in pre-season forecasts of Fraser sockeye returns has eroded among fisheries managers and harvesters in recent years because observed returns have frequently diverged outside of the 80% prediction interval (i.e., <10 or >90 of posterior quantiles in Bayesian statistical approaches; DFO 2021). Large differences between pre-season forecasts and observed returns have led to potentially reduced fishery opportunities, missed management targets and escapement goals, as well as increased conservation concerns for Fraser sockeye salmon stocks. Therefore, it is imperative to improve forecasting ability for Fraser sockeye; in particular, it has become necessary to incorporate new information derived from recent scientific research into the forecast process such that concurrent climate change is reflected in the forecast models.

Sockeye returns in any given year are influenced by the abundance of their parental spawners (i.e, escapement or sockeye returns that reach the spawning grounds), the proportions of age classes, and the survival rate of the adult recruits across the entire life cycle from egg to adult (Grant et al. 2010; Hawkshaw et al. 2020a, b). A wide variety of forecast models are typically available at a stock level for Fraser sockeye from non-parametric ‘naïve’ models (i.e., models using recent and historical adult recruits only without considering spawners) to complex stock-recruitment models that incorporate environmental variables into the relationships between spawners and subsequent recruitment (Cass et al. 2006; Grant et al. 2011; DFO 2018; Hawkshaw et al. 2020a, b). More advanced forecast methods have also been developed that use Bayesian statistical approaches to assess uncertainties associated with abundance estimates (Cass et al. 2006; Grant et al. 2011; Akenhead et al. 2016a). Model selection for Fraser sockeye forecasts has been an expert-driven process, primarily adopting forecast models identified based on Jack-knife cross validation (i.e., analysis being carried out by randomly leaving out one year’s data at a time) conducted for the period of 1997 to 2004 (Grant et al. 2011).

Since the 1990s, Fraser sockeye has generally declined with several populations being assessed as endangered by the Committee on the Status of Endangered Wildlife in Canada (COSEWIC; COSEWIC 2017). Forecast models that were selected by Grant et al. (2011) more than a decade ago have been continuously used in pre-season forecasts despite the fact that their explanatory power was found low (Akenhead et al. 2016a; DFO 2021). Moreover, environmental covariates such as coastal sea surface temperature (SST) and the Pacific Decadal Oscillation (PDO) that were shown to be effective in reducing the degree of unexplained stock-recruitment variation in the past two decades (Cass et al. 2006), are no longer applicable for most stocks (Litzow et al. 2020). These time-varying effects of environmental conditions on population and community processes, i.e., non-stationary relationships, are becoming more prevalent in the face of climate change, and have been increasingly recognized (e.g., Litzow et al. 2018, 2020; Ohlberger et al. 2022). Such non-stationarity requires that previously established Fraser sockeye forecast models be continuously evaluated through annual retrospective analysis before carrying out pre-season forecasts. Furthermore, external drivers that are potentially responsible for the interannual and long-term variability of sockeye abundance and productivity continue to be explored and characterized in a changing environment (McKinnell 2008; Ruggerone et al. 2021).

In this study, we developed a framework that can be easily used by fisheries managers or other stakeholders to evaluate pre-season forecast models and to identify external drivers of importance for forecasting Fraser sockeye salmon returns on an annual basis. Specially, we developed computer codes to conduct annual retrospective analysis of multiple Fraser sockeye forecast models and for the first time employ Taylor diagrams (Taylor 2001) with good visualization to display simultaneously three statistics of predictive power, including Pearson’s correlation coefficient (denoted as correlation), normalized standard deviation and normalized root-mean-square error (RMSE). Normalization is calculated as dividing both the RMSE difference and the standard deviation of model prediction for a set of years by the standard deviation of observed abundance for the same set of years. The Taylor diagram is a useful visualization tool that has been employed to evaluate the performance of a variety of models such as those for ocean circulation (Lamine et al. 2022), satellite derived chlorophyll a measurements (Lee et al. 2015), biogeochemical dynamics (Salihoglu et al. 2017), and stock assessment and population dynamics of tuna species (Kell et al. 2016; Inna et al. 2019). The framework established through this study can be directly implemented in the process of Fraser sockeye forecast so as to improve the performance of sockeye forecasts and assist stock assessment and fisheries management planning.

1. **Materials and methods**

**2.1 Fraser sockeye spawner and recruit data**

In this study, we used the same spawner and recruitment data of Fraser sockeye stocks within the Fraser River watershed (Figure 1) that were approved for the 2022 pre-season forecast (DFO, in press). The spawner data have been collected and maintained by DFO since 1938 (Grant et al. 2011). The recruitment data have been compiled by Pacific Salmon Commission (PSC) since 1952. Similar to Grant et al. 2010, in this study, we treated the recruitment data as observation to represent the true abundance of sockeye returns to the Fraser River mouth. The details on data quality are available in Ogden et al. (2015). Based on shared timing of migration and their entry into the Fraser River, the stocks are aggregated into four management units: Early Stuart, Early Summer, Summer, and Late Run. In this study, we focused on 18 ‘lake-type’ sockeye stocks that typically spend two years in a nursery lake before migrating seaward and return to their natal rivers at age 4 or 5.

For all stocks, we chose female spawners from the 1948 to 2018 brood years as an indicator of spawner abundance (also called effective female spawners). Effective female spawners (EFS) were defined as “the product of the total female spawners multiplied by female spawner success, where female spawner success was calculated as the proportion of eggs (0%, 50%, or 100%) successfully spawned, based on spawning ground carcass surveys” (Grant et al. 2011; Ogden, et al. 2015). Recruits included both sexes with the most recent recruitment data (4- and 5-year-old) available up to the 2017 brood year (i.e., the primary brood year returning in 2021). For the Cultus Lake sockeye stock, juvenile abundance estimated during outmigration at a counting fence instead of spawner abundance was used to build the stock-recruitment relationship, because the quality of the adult escapement data is variable and they are heavily affected by enhancement activities (Ackerman et al. 2014).

Similar to Grant et al. 2010, we assumed age composition of Fraser sockeye adults generally followed a similar pattern among brood years around the historical average, which was typically dominated by age 4 year class. For example, to get the total recruits of 2022, we used 2018 brood year EFS to produce a total recruits multiplied by the average stock-specific proportion of age-4 recruits, and used 2017 brood year EFS to produce a total recruits multiplied by the average stock-specific proportion of age-5 recruits. Ten of the eighteen stocks had time series of paired spawner and recruit estimates from brood years going back to 1948, and the remaining eight stocks had shorter time series (Nadina: 1973; Gates: 1968; Scotch: 1980; Fennell: 1967; Weaver: 1966; Portage: 1953), depending on the availability and quality of the data as well as consistency with previous pre-season forecasts (DFO 2018).

* 1. **Environmental and ecological covariates**

Previous pre-season forecasts have incorporated a few local and large-scale environmental covariates in the Northeast Pacific Ocean, including (1) PDO in the winter preceding outmigration (November-March, Mantua et al. 1997), (2) monthly average SST (April-June) from Entrance Island lighthouse (Ei.SST, Strait of Georgia, near Nanaimo, BC, Canada), (3) monthly average SST (April-July) from Pine Island (Pi.SST, Northeast corner of Vancouver Island) of the year of outmigration, (4) peak [Fraser River discharge (FRD.peak), and (5) average Fraser River discharge (FRD.mean) from April to June of the outmigration year, both measured at Hope, BC, Canada](https://wateroffice.ec.gc.ca/) (for data sources please refer to DFO 2018; Hawkshaw et al. 2020a, b). All time series of environmental indices were aligned with smolt outmigration year because that is when they are most influential on marine survival of sockeye and other Pacific salmon species (Cass et al. 2006).

Exploration of new biological and environmental covariates that may explain the interannual variability in Fraser sockeye recruitment has long been suggested (Grant et al. 2010; MacDonald and Grant 2012; DFO 2015; Hawkshaw et al. 2020a, b). Recent analyses found that survival of Fraser sockeye exhibited a similar temporal pattern among all stocks, indicating that Fraser sockeye stocks might be related to oceanic environmental conditions and competition among Pacific salmon at a global scale (Akenhead et al. 2016a; Connors et al. 2020; DFO 2021; Ruggerone et al. 2021; Rosengard et al. 2021). Therefore, we included Gulf of Alaska SST (GOA.SST) time series, which was extracted from the COBE SST database (<https://psl.noaa.gov/data/gridded/data.cobe.html>), matching with the Fraser sockeye distribution areas identified based on tagging studies (Myers et al. 1996; Ishii et al. 2005; Supplementary Figure S1). We also included the total abundance (i.e., catch plus escapement) of Pacific salmon returning from the North Pacific Ocean to streams in Asia and North America as covariates in the stock-recruitment models. Specifically, we chose annual abundance of adult pink, chum, and sockeye stocks, respectively, and the sum of all three salmon species abundance (Salmon.Total) from 1950 – 2020 (brood year 1948 – 2018) as covariates as ocean entry year is believed to be the most important time for sockeye marine survival (Ruggerone and Irvine 2018; Ruggerone et al. 2021). Altogether, ten covariates (six environmental and four ecological) were included in the pre-season forecast models (Figure 2), hypothesizing that inclusion of these covariates can help account for more of the environmental effects on interannual variability in the survival of all Fraser sockeye stocks.

**2.3 Forecast models**

In this study, we re-evaluated the previously explored models, including eleven naïve and six biological models (Table 1 and Table S1). The biological models (or Stock-Recruit models) of Ricker, Power, and Larkin were also applied to data of dominant years only (i.e., RickerCyc, PowerCyc, and LarkinCyc) for the purpose of understanding whether or not the dominant year class was driven by different biological and environmental processes. With the Ricker and Power models, we also coupled them with each of the ten covariates to form ten new models. For model descriptions of the 37 pre-season forecast models, please refer to Table 1 and Table S1. Following previous forecasts (e.g., Cass et al. 2006; Grant et al. 2011; MacDonald and Grant 2012; DFO 2018; DFO in press), we used a Bayesian statistical approach to estimate biological model parameters and MCMC (Markov Chain Monte Carlo) to assess estimation uncertainties. Model specifications are documented in Supplement tables (Priors.csv and mcmc\_specs.csv).

**2.4 Retrospective analysis and model evaluation**

To evaluate the 37 forecast models for the 18 major stocks, we conducted a one-step-ahead retrospective analysis (i.e., iteratively stepping forward through time as each step added a new year to the estimation dataset) for the period between 2009 and 2020. For each stock, all 37 forecast models were ranked using each of the following four performance measures related to forecast error: mean absolute error (MAE), root-mean-square error (RMSE), mean absolute relative error (MARE), and mean absolute proportional error (MAPE). MAE is the sum of absolute errors between forecast (*y*) and observation (*x*) divided by the time series length *T*:  (Willmott and Matsuura 2009). RMSE provides a measure of forecast error variance reflecting both bias and uncertainty: (Hyndman et al. 2006). MARE () and MAPE ( ) reflect the long-term bias of forecasts. For each model, ranks across the four performance measures were averaged to generate an overall rank. We then calculated relative rank based on the overall rank to make comparisons across stocks: . We also provided individual relative rank for RMSE, MAE, MARE and MAPE, and RMSE has been commonly used by fisheries managers and other similar studies (Haeseker et al. 2008; Ovando et al. 2022).

We used Taylor diagrams as a visualization tool to guide the process of model selection both in terms of accuracy and uncertainty. The correlation (with values ranging from -1 to 1) between forecasts hindcasted by each model for the years 2009 – 2020 and the observed returns represents forecast accuracy. Correlation values greater than 0.75 indicate high forecast accuracy while smaller values particularly negative values imply inaccurate forecast. The standard deviation represents the magnitude of the forecasts variation relative to the observation variation for the given years. The observed sockeye returns from 2009 – 2020 (about 3 cycles for some major stocks; Supplementary Figure S2) were normalized as a reference point (i.e., observation). The overall performance of a model was expressed as a relative position on the diagram to the observation. Generally, a model that has relatively high correlation, small RMSE, and similar standard deviation as observations is desired. A perfect model is located at the reference point with both correlation andstandard deviation being 1 and RMSE being 0. All models, analyses and visualization were programmed using R 4.2.0 (R Core Team 2022).

1. **Results**

**3.1 Forecast model comparisons**

Relative rank based on overall rank of the 37 forecast models across the 18 stocks showed clear patterns of generally poor performances by the naïve models but relatively good performances by either Ricker or Power models coupled with the newly included covariates (i.e., GOA.SST, Sockeye, Chum, Pink and Salmon.Total) (Figure 3). Relative rank based on RMSE, MAE, MARE, and MAPE showed similar patterns to those based on overall rank (Supplementary Figure S3 a-d).

Correlations between model forecasts and observations were generally lower for the naïve models than biological models (Figure 4). Correlations were more homogeneous within each stock (across all models) than across the stocks (using the same model). This finding may indicate that model accuracy can be stock-dependent as a consequence of either data quality or specific stock dynamics. The small stocks of Gates and Raft had particularly low correlations across all models, possibly associated with inaccurate yet more economical survey methods applied to insignificant stocks (DFO, 2021). Similarly, standard deviation revealed the disadvantage of the naïve models, which tend to be farther away from 1 (Supplementary Figure S4). Interestingly, Naïve models (in particular, LLY model) perform the best among others for these stocks with the correlation 0.3-0.5. Standard deviation also showed a high degree of consistency across all the models other than the naïve models, particularly for the three stocks, Nadina, Stellako, and Late Stuart, for which the model forecasts either didn’t capture uncertainty or had inflated variability compared to observations.

For each of the 18 Fraser sockeye stocks, we presented the retrospective hindcast results (for the period 2009 – 2020) of the 37 forecast models on a Taylor diagram relative to the time series of observations (shown as a black solid circle; Figure 5a,b). The historical forecast based on previously selected models was also presented (shown as a black solid square; Figure 5a, b) relative to the observation. In general, the historical forecast performed poorly compared to many other models. Correlations between the historical forecast and the observation were less than 0.5 for half of the stocks with some close to or even below zero. For all stocks, there was at least one model that outperformed the historical forecast, indicating that had we chosen a good performing model consistently throughout the years, we would have had more accurate forecasts compared to those obtained from the forecast models historically chosen.

Among biological models, the Power model coupled with new covariates (blue icons in Figure 5) generally outperformed other models with higher correlations, similar standard deviations to those in observations, and smaller RMSEs (Figure 5a, b). The performance of the Power model combined with the previously adopted covariates (green icons in Figure 5) was similar to the best models, yet they were rarely used in previous forecasts (Grant et al. 2011; DFO 2018; Hawkshaw et al. 2020a, b). The Ricker model coupled with new covariates (red icons in Figure 5) also seemed to be better compared to those with old covariates (pink icons in Figure 5) with the latter having been frequently used in the previous forecasts. For most stocks, the best model forecasts based on the three statistics (correlation, standard deviation, and RMSE) shown in the Taylor diagrams were in good agreement with the observation, reaching correlations between 0.75-0.95. However, for a few stocks such as Gates and Raft, none of the existing models produced an accurate forecast, probably due to data quality issues as aforementioned. Naïve models (purple icons in Figure 5) tended to have larger standard deviations and performed worse than biological models in general (except LLY for some stocks).

Age-specific Taylor diagrams are presented for all 37 forecast models relative to the historical forecast (black solid square in Figures S5-6) and the observation (black solid circle in Figures S5-6) for age 4 and age 5 returns, respectively (Supplementary Figure S5 and S6). The age-4 Taylor diagrams showed similar results as the diagrams produced for total abundance (Figure 5a,b), since age 4 was the dominant age class for most stocks. Power models coupled with new covariates consistently outperformed other models for many stocks. The historical forecast of age 4 showed similarly poor performance for most stocks (Supplementary Figure S5). For age 5, sibling models (yellow squares in Figures S6) performed better for many stocks than the naïve and biological models (Supplementary Figure S6).

**3.2 Model selection**

In 2021, when sockeye forecasts was conducted using previously selected forecast models, the forecast error (defined as ) was 92%, being the highest among all the sockeye runs in Northeast Pacific (Supplementary Table S2). In 2022, Taylor diagrams started to be implemented to identify the best performing model for many Fraser stocks, as a result, the forecast error has reduced to 30%, much lower than some of the other sockeye runs (Supplementary Table S2). This reduced forecast error also represents a significant improvement compared to the past few years (Hawkshaw et al. 2020a, b). In 2023, when Taylor diagrams were further implemented, the preliminary results showed that the forecast error was -3% only (pre-season forecast is 1.564 million, in-season adopted run size is 1.606 million; PSC, unpublished document Sep 19, 2023).

To understand how the models that were ranked first based on either overall rank or one of the three statistics (correlation, standard deviation, and RMSE) performed compared to other models, we examined the largest three stocks (Chilko, Late Shuswap, and Quesnel) as examples (Figure 6). For Late Shuswap, where the 2022 forecast was done by the historically selected model (RickerEi.SST), the forecast resulted in a deviation of -57%, equivalent to a difference of 1.9 million sockeye salmon relative to the observed returns. In contrast, the model ranked first based on overall rank (i.e., RickerGOA.SST) resulted in the most accurate forecast with the smallest uncertainty (Figure 6). The RickerPi.SST model also performed relatively well, being best according to either correlation, standard deviation, or RMSE. The RS1 model, ranked third based on standard deviation following RickerPi.SST and RickerGOA.SST models, also performed better than the RickerEi.SST model, which may indicate that standard deviation is an important factor to consider in the model selection process. The importance of standard deviation can also be seen with Chilko and Quesnel, where the top-rank models based on standard deviation (i.e., R1C and RickerSockeye, respectively) also performed the best with the highest accuracy and lowest uncertainty (Figure 6). By contrast, RickerCyc, although highest ranked according to correlation and RMSE, performed poorly for Chilko. Similarly, the Larkin model performed poorly for Quesnel despite being ranked highest according to correlation.

1. **Discussion**

Forecasting Fraser sockeye salmon returns in the upcoming year is a challenging task given the complex life history and the dynamic freshwater and marine habitats in which they reside. In recent years, Fraser sockeye forecast accuracy has been declining, frequently going below 10 or above 90 posterior quantiles, i.e., at the extremes ends of the 80% prediction interval (DFO 2021). For example, in 2021, the median forecast for the total Fraser River sockeye return was 1.33 million fish while the observed return was 2.549 million, 92% above the forecast (Supplementary Table S2). Despite these kinds of extreme differences, the official forecast is widely used by various groups for important planning decisions, including escapement survey planning, fishery opening and licensing, allocation among commercial, recreational and indigenous fisheries, hatchery enhancement activities, research and development, and international salmon treaty negotiation (Haeseker et al. 2008; Michielsens and Cave 2019; Hawkshaw et al. 2020a, b; DFO in press). Developing a framework to improve and streamline the existing forecast process is highly desired for a broad range of stakeholders. In this paper, we developed a framework that allowed fisheries managers and other stakeholders to evaluate the performances of pre-season forecast models on an annual basis by using Taylor diagrams, and to identify external drivers of importance for forecasting Fraser sockeye returns. Specifically, for the first time we incorporated new environmental and ecological covariates (i.e., GOA.SST, Pink, Chum, Sockeye, Salmon.Total) into the Ricker and Power forecast models, which has resulted in consistently robust forecasts across all 18 Fraser sockeye stocks, regardless of data quality and survey method changes temporally over multiple decades and spatially over the entire Fraser watershed.

**4.1 Moving towards ecosystem-based fisheries management**

Sockeye salmon have a complex life history, going through vastly different freshwater and marine environments at different life stages, resulting in very unpredictable impacts on stock productivity (e.g., Ricker 1997; Akenhead et al. 2016b; Huang et al. 2021). It can be helpful to have reliable modelling of their complex life history and particularly of the impacts of climate change, ecosystem phase transitions, and competition for food and other resources when conducting forecast of sockeye returns (e.g., Patterson et al. 2016; Satterthwaite et al. 2020; Connors et al. 2020; Wainwright 2021; Ohlberger et al. 2022; Kaeriyama 2022). Moving pre-season sockeye forecast and management towards ecosystem-based approaches that account for ecosystem processes, including fishing and/or climate variability in conjunction with species interactions (Sissenwine and Murawski 2004; Link 2011) could improve the performance of forecast models and by extension, the management systems they serve. Multiple studies have found that sockeye salmon in BC have been negatively affected by the significant increase of Pink salmon abundance in the north Pacific Ocean in recent decades through competition for limited resources on both broad and localized scales (Ruggerone and Connors 2015; Connors et al. 2020; Ruggerone et al. 2021; Litz et al. 2021). Even though existing Fraser sockeye stock assessment and fisheries management are still heavily single-species based, our study is a step forward towards ecosystem-based fisheries management by showing that incorporating the covariates of pink, chum, other sockeye abundance, and their combined abundance, implicitly accounting for inter-specific and intra-specific competition, will improve forecast performance. Our forecast framework revealed that models combined with the salmon covariates were generally highly ranked retrospectively from 2009-2020, highlighting that Fraser sockeye dynamics were closely related to the abundances of other Pacific salmon stocks in the north Pacific Ocean. In addition, the current area was based on previous tagging studies on distribution of sockeye salmon (Myers et al. 1996; Ishii et al. 2005; Supplementary Figure S1), however, recent high-seas surveys (2019-2020, 2022) in the Gulf of Alaska suggested that Fraser sockeye distribution could be even wider in the North Pacific Ocean, with one sockeye caught in the eastern hemisphere (176.2˚E). These surveys also found that Fraser sockeye shared marine habitat with sockeye salmon from other river systems in cooler waters as well as Pink and Chum salmon, although the spatial overlap with the latter two species were less intensive (https://www.youtube.com/watch?v=thCmnoPEw6s). The extended period of residence in the GOA by Pacific salmon, their high degree of spatial overlap in this region, and evident co-variability among salmon species mean that the GOA is an area of significance for salmon abundance.

The GOA is influenced by decadal thermal variability and has experienced three marine heatwave events since 2013, resulting in unprecedented reductions in fishery recruitment and shifts in the biological community of this region (Blaisdell et al. 2021). The inclusion of GOA.SST into the pre-season sockeye forecast was another step of moving forward towards ecosystem-based fisheries management. The RickerGOA.SST model (Ricker coupled with GOA.SST) performed extremely well for the Late Shuswap stock (one of the three major Fraser sockeye stocks) compared to all other models with the exception of RickerPi.SST (Figure 3). Had the RickerGOA.SST model been adopted in 2022 for the Late Shuswap stock, the median forecast for this stock would have been reduced from 3.42 million to 1.48 million (equal to the observation), and the error for Fraser River sockeye would have been reduced from -30 to -10% (Supplementary Table S2). It is also worth noting that the current GOA.SST map only represents part of the northeastern Pacific Ocean (Supplementary Figure S1), which can be expanded when we understand the Fraser Sockeye distribution in GOA better. Nevertheless, our paper sheds some light on how forecasts can be improved by incorporating biotic and abiotic metrics from ecosystem perspectives. Overall, we showed that the forecast models coupled with newly added covariates not only reduced the differences between forecast and observation but also reduced uncertainties, which would be a great help to fisheries managers who tend to face high pressure from political parties when forecast is inaccurate or has high uncertainty.

**4.2 Taylor diagram and future implications**

The Taylor diagram allowed us to visually compare time series of forecasts from 37 forecast models along with the historical forecasts against the observations in one figure both in terms of forecast accuracy and uncertainty. This approach can help simplify the model selection process by illustrating performance quantitatively among all forecast models. Based on the Taylor diagrams for all 18 Fraser sockeye stocks, we concluded that historically selected forecast models were not ideal compared to many other forecast models. The framework developed through this study can be adopted for future model selection and forecast processes (DFO, in press). For the first time, standard deviation comparison between model prediction and observation has been taken into consideration in salmon forecast. This information could be helpful to resource managers when making management decisions. For example, a model with larger SD than observation could result in larger uncertainties during dominant years, but may be less for non-dominant years. Although certain forecast models were chosen because of their overall performance comparing to other models, managers should be aware of their caveats if their SD are too large or too small comparing to SD of observations.

Many Fraser sockeye stocks exhibit strong cyclic patterns with a dominant year every four years (Supplementary Figure S2). Forecasting for the dominant years was more challenging and was usually accomplished with lower accuracy compared to other years (DFO, 2021). Our retrospective analysis only covered 12 years (3 cycles); however, as the time series extends in future, it will be interesting to separate dominant years and plot cycle-specific Taylor diagrams to identify models specifically tailored to dominant years. It will also be interesting to examine how the relative positions of these forecast models evolve over time. Theoretically, informative models will move closer to the observation as more years of data become available, assuming that the observation is a reliable reference. Models with false alarms (bad model but appear to have good performance temporarily) will either wander around or move further away from the observation. Monitoring the moving directions of each model for multiple years may provide insights into their relative forecasting skill or potential for improvement, based on relative differences compared to observations.

Currently, our framework allows us to select a model by visually looking at the models’ relative positions in Taylor diagrams, which could be subjective. For example, a best RMSE model may not have best standard deviation or best correlation or vice versa. To be more objective, a single distance metric, similar to Mohn’s rho value (Mohn, 1999), can be developed with proper weights among three metrics and integrated to automate the model selection process to come up with best machine-suggested models.

**4.3 Forecast models for non-stationary relationships**

It has been commonly recognized that correlations established based on past observations can be easily broken down due to the non-stationary nature of environmental changes (e.g., Peterman & Dorner 2012; Malick 2020; Litzow et al. 2020). Such non-stationarity calls for continuous evaluation of previously established forecast models. While annually evaluating the performance of forecast models that include environmental variables may facilitate more accurate and precise forecasts, actually incorporating the best performing models into the management process every year can be disruptive and may risk chasing noise (Winship et al., 2015). Therefore, we should be cautious interpreting results based on single-year performance but rather focus on performance stability across multiple years if such risk was introduced by model overfitting or data quality issues. Winship et al. (2015) identified a lot of inter-annual variation in the environmental covariates for inclusion in a forecast model for a Chinook salmon stock. When using a moving data window or down-weighting more distant data, the performance of various models changed even more frequent over time and no stable specific functional relationship was identified in their study. There is a need to conduct more research to identify models that perform more consistently over time and more robustly to environmental changes, particularly in the light of the connections between life history characteristics of different taxa and forecast model performances (Ward et al. 2015). In this study, we also found that sibling models outperformed others for age-5 for most of the Fraser stocks, suggesting that sometimes simple model can provide satisfactory forecast without knowing exact mechanisms. This is consistent with other studies (Scheuerell and Williams, 2005). However, there may be even a greater need to develop management systems that are robust to forecast uncertainties (Wainwright 2021), as improving forecast skills solely not necessarily implies improving fisheries management (Rupp et al. 2012).

For the Chilko and Quesnel sockeye stocks, the RickerCyc and Larkin models performed poorly although each model was ranked first based on correlation. Based on this result, we infer that the correlations established for the period 2009 – 2020 may be deteriorating as environmental conditions have been constantly changing. On the other hand, we found that the top-rank models based on standard deviations performed the best with the highest accuracy and lowest uncertainty for the two major stocks Chilko and Quesnel (Figure 6). With the Lake Shuswap sockeye stock, all three models with the first three ranks based on standard deviation (i.e., RickerPi.SST, RickerGOA.SST, and R1C) performed better than the historical forecast (Figure 6). These findings may show that standard deviations established based on past observations are not as easily broken down as correlations and thus could be considered as an important factor in the model selection process.

In terms of stock-recruitment relationships, parametric models, such as Ricker, Power, and Larkin, have been primarily employed in the past (Ricker 1997; Cass et al. 2006; Grant et al. 2010, 2011; Akenhead, et al., 2016a; DFO 2018; Hawkshaw et al. 2020a, b). However, these parametric models have limited forms to encompass the inherent uncertainties associated with large annual variability in multiple environmental factors that affect salmon dynamics throughout their complex life history stages (e.g., Healey, 2011; Martins et al., 2012; Padilla et al., 2015). They are thus inadequate to incorporate anthropogenic changes (Akenhead et al., 2016b), which can result in unexplained residual patterns in the stock-recruit relationships, resulting in less accurate pre-season forecasts. In addition, these parametric models are also difficult to deal with non-stationary sockeye dynamics (Peterman & Dorner 2012; Malick 2020), which may contribute the Chilko sockeye return to have been persistently over-estimated in the last decade (DFO 2021). Climate-induced non-stationarity in relationships between marine environments and fish communities has been increasingly recognized (e.g., Litzow et al. 2018, 2020; Ohlberger et al. 2022) and it has been advocated that new approaches capable of dealing with non-stationarity in sockeye dynamics should be explored (Peterman & Dorner 2012; Malick 2020).

Alternative approaches such as generalized additive models (Wood 2017), boosted regression trees (Elith et al. 2008), random forest (Breiman 2001), empirical dynamic models (Ye et al. 2015), and artificial neural network (Ripley, 1996), use algorithms to learn the relationships between responses and predictors and are not confined to certain functional forms. These more flexible models can be developed to represent non-stationary stock-recruitment relationships in the dynamics of sockeye stocks. In addition, an ensemble modeling approach has been developed in recent years (Ovando et al. 2022) and it could be useful as it could take advantages of multiple models and improve forecast skills. Recently discovered similar temporal patterns of Fraser sockeye survival among all stocks (Akenhead et al. 2016a; Connors et al. 2020; DFO 2021; Ruggerone et al. 2021; Rosengard et al. 2021) also suggest that methods based on shared trends and autocorrelation can be explored. Such methods have proven effective in other situations (e.g. Winship et al. 2015; Johnson et al. 2016; DeFilippo et al. 2021), while environmental relationships can be broken down, resulting in higher forecast errors particularly for highly variable marine fish stocks (e.g. Ward et al. 2014). With all alternative approaches being incorporated, Taylor diagrams can then be applied to study how these models perform compared to the traditional naïve models and parametric models. In the future, we can also incorporate multiple covariates at the same time to evaluate their relative importance in forecasting sockeye return. In summary, we advocate for continuously evaluating all potential forecast models in the face of increasing environmental change and the development of new models able to deal with non-stationary relationships between environment and sockeye dynamics. This process of continuous evaluation should lead to the identification of good performing forecast models and to improved understanding of the impacts of environmental and ecological factors on the performance of Fraser sockeye forecast.

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**Data availability**

All data and code are available in a Github repository. <https://github.com/yi-xu/Sockeye_paper> Access can be granted when contacting corresponding author.

**Competing interests**

The authors declare there are no competing interests.

**Supplementary material**

Supplementary data are available with the article at https:

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**Figure captions**

Figure 1. Locations of 18 major Fraser Sockeye salmon stocks where spawning data were collected. Color indicates different run type: red for Early Stuart, green for Early Summer run (including Bowron, Fennel (Upper Barriere), Scotch, Nadina, Pitt, Seymour, and Gates), blue for Summer run (including Chilko, Quesnel, Late Stuart, Stellako, and Raft), and purple for Late run (including Late Shuswap, Cultus, Portage, Weaver, and Birkenhead).

Figure 2. Interannual variations of ten environmental and ecological variables from 1950-2020, including sea surface temperature (SST, unit in oC) at Entrance Island and Pine Island, mean and peak Fraser River discharge at Hope (unit m3/s), Pacific Decadal Oscillation (PDO), SST in the Gulf of Alaska, North Pacific Sockeye, Chum, Pink salmon abundance (catch plus escapement, number in million), and combined total of the three salmon species.

Figure 3. Relative rank among all 37 forecast models for all 18 stocks. The relative rank (scale of 0-1) for an individual stock was derived from the overall rank table by dividing the rank of a model by the number of models evaluated for this specific stock. Blank indicates the model was not applicable to the stock.

Figure 4. Correlations between retrospective forecasts for the period of 2009 – 2020 based on all 37 forecast models and observations for all 18 Fraser Sockeye stocks. Correlation between historical forecasts (denoted as Forecast) and observations was also shown.

Figure 5. (a) Taylor diagrams for Early Stuart Sockeye stock, seven Early Summer run (Bowron, Fennel (Upper Barriere), Scotch, Nadina, Pitt, Seymour, Gates), and one Summer run (Chilko). (b) Taylor diagrams for four Summer run (Quesnel, Late Stuart, Stellako, Raft) and five Late run (Late Shuswap, Cultus, Portage, Weaver and Birkenhead). Each Taylor diagram compares 37 model forecasts and the historical Forecast (black solid square) against the Observation (black solid cycle on the x-axis). The distance from the origin is the normalized standard deviation with the normalized value for observations being 1. The angle describes the correlation (0 to 1) between model forecasts and observations. The dashed arcs around the Observation illustrate the root-mean-square error (RMSE). Models with negative correlations are not shown for each stock. The closer the model is to the Observation, the better predictive power the model has.

Figure 6. Comparisons among the forecast based on the historically selected model (Forecast) for the year 2022 and the forecasts by three top-ranked models using either best overall rank (Rank), standard deviation (SD), correlation (R), or root-mean-square error (RMSE) for three most abundant Fraser sockeye stocks: Chilko, Late Shuswap and Quesnel. The 50 percentile (P50: solid black line) and forecast distribution (box: P25-P75, range: P10-P90) were compared with the observation of the Sockeye return in 2022 (PSC preliminary results). \*The Chilko stock used age-specific forecast models with age 4 using RickerEi while age 5 Sibling; Descriptions for the model abbreviations are in Table 1.

Table 1. Descriptions of 37 models under three categories (A: non-parametric/naïve, B: biological, and C: biological models coupled with covariates) along with Sibling model that was used as an additional option for forecasting age 5 specifically, where *R* is recruitment, *S* is spawner, *a* for the intercept measuring density-independent per capita *R*, and *b* for the slope measuring density-dependent *R*.

| **MODEL CATEGORY** | **DESCRIPTION** |
| --- | --- |
| **A. Non-Parametric (Naïve) Models** | |
| LLY, R1C, R2C, RAC, TSA, RS1, RS2, RS4yr, RS8yr, MRS, RSC | Models only used returns from selected years, not considering spawners. Detailed model descriptions are listed in Supplementary Table S2. |
| **B. Biological Models (Stock-recruit Models)** | |
| Ricker | Bayesian Ricker model, loge(*Rt*/ *St*) = *a* – *b* *St* + *Ɛt* |
| RickerCyc | Same as above, using cycle line data only |
| Power | Bayesian power model, loge(*Rt*) = *a* + *b* loge (*St*) + *Ɛt* |
| PowerCyc | Same as above, using dominant year data only |
| Larkin | Bayesian Larkin model, loge(*Rt*) = *a* + *b*1 loge (*St*) + *b*2 loge (*St*) + *b*3 loge (*St*) + *Ɛt* |
| LarkinCyc | Same as above, using cycle line data only |
| **C. Biological Models** (B for Ricker or Power) **coupled with Environmental and Ecological Covariates** | |
| B\_FRD.mean | B coupled with Mean Fraser discharge flow from April to June |
| B\_FRD.peak | B coupled with Peak Fraser Discharge at a given year |
| B\_Ei.SST | B coupled with Mean Entrance Island sea-surface temperature (SST) from April to July |
| B\_Pi.SST | B coupled with Mean Pine Island SST from May to July |
| B\_PDO | B coupled with Pacific Decadal Oscillation in winter preceding outmigration from November to March |
| B\_GOA.SST | B coupled with Mean Gulf of Alaska annual SST |
| B\_Pink | B coupled with Abundance of pink salmon in the North Pacific Ocean |
| B\_Chum | B coupled with Abundance of chum salmon in the North Pacific Ocean |
| B\_Sockeye | B coupled with Abundance of sockeye salmon in the North Pacific Ocean |
| B\_Salmon.Total | B coupled with Abundance of pink, chum, sockeye salmon altogether in the North Pacific Ocean |
| **Additional model for age 5** | |
| Sibling | Bayesian sibling model, loge(*R5, t*) = *a* + *b* loge (*R4, t-1*) + *Ɛt* |

Figure 1.Map

Description automatically generated

Figure 2.

A picture containing timeline

Description automatically generated

Figure 3.

A picture containing background pattern

Description automatically generated

Figure 4.

Background pattern

Description automatically generated

Figure 5a

Chart, map

Description automatically generated

Figure 5b.

Map

Description automatically generated

Figure 6.

Chart, box and whisker chart

Description automatically generated